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Chapter 3

Biodiversity and Host–Parasite (Co)Extinction



Jeroen van Dijk and Kenneth De Baets

Abstract Parasitism is one of the most common modes of life, and yet it is often disregarded or ignored in nature conservation. We are at the brink of the sixth mass extinction and in order to assess the extinction risk of both parasites and their hosts, we first need to fully understand the role and function of parasites in ecosystems. Parasites might play an active role in their host’s extinction, and coextinction has been postulated to be the most common mode of extinction. However, parasites may be able to survive their host’s extinction through host switching, perhaps to a more abundant host, for example. The dilution effect has been described as an important natural defense mechanism for the host: higher biodiversity is associated with lower infection risk. Discussed here is the importance of biodiversity and host–parasite associations and (co)extinction, and the role the fossil record has in filling the knowledge gap regarding deep-time host–parasite interactions.

Keywords Biodiversity · Co-extinction · Dilution effect · Host–parasite associations · Fossil record · Host switching

3.1 Introduction

We are currently witnessing the sixth mass extinction or at least its beginning, as became clear in the last decades (Wake and Vredenburg 2008; Barnosky et al. 2011; Payne et al. 2016). The Earth’s biodiversity is decreasing, with more and more species disappearing. Even more worrying though is that a large portion of the estimated number of existing species today are still awaiting description (Mora et al. 2011). There is some bias in human pity: it seems dramatic when wild cats and rhinoceroses disappear, but who will cry for mites, fleas, and lice? Costello et al.

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(2013) famously asked whether we can name all species before they go extinct. The question raised is not trivial: Is naming every single species necessary, or is the cessation of some groups a loss that we can easily bear? Parasites, for example, are more or less considered species *non gratae* in nature conservation (Strona 2015; Dougherty et al. 2016; Carlson et al. 2020).

The macroevolutionary point of view defends parasites, arguing that they might fulfill an important stabilizing function for maintaining the balance of ecosystems (Seilacher et al. 2007; Lafferty 2012), a balance that might shift during mass extinctions (Seilacher et al. 2007). Dougherty et al. (2016) and Carlson et al. (2020) suggested to include parasites into nature conservation, but this requires a framework that tests for the viability of host–parasite assemblages within a population in order to assess extinction risk. And here is where we come to some quite important and interesting questions: How important are parasites to an ecosystem (beyond being a nuisance to their hosts)? And, if they are important, what is their role? More importantly, how does parasite diversity and extinction relate to host diversity and what happens to parasite diversity during mass extinctions?

3.2 Host–Parasite Biodiversity

Parasitism is one of the most common modes of life on Earth (Poulin and Morand 2000), which has independently evolved at least 223 times within metazoa (Weinstein and Kuris 2016), but also throughout the entire tree of life (Bass et al. 2015). Parasites can be highly abundant in ecosystems, and not only reach high diversity levels, but also take up a substantial part of an ecosystem’s biomass (Kuris et al. 2008). Indeed, ecosystem functioning seems to improve with increased diversity of parasite species (Hudson et al. 2006).

Nonetheless, parasites are often seen as disgusting, associated with diseases and low hygienic conditions, and are therefore probably the most ignored group of organisms in nature conservation (see Dougherty et al. 2016). When it comes to protecting the natural world, parasites are often overlooked (e.g., biodiversity counts), because they do not come to mind (e.g., too small), or because they simply lack charisma (Dunn et al. 2009). Instead, they are often seen as part of the problem: a threat to wildlife, and should therefore be eradicated. Medical and veterinary sciences attempt to remove parasites from both human and animals by all means necessary, as evidenced by the tragic fate of the condor louse (*Colpocephalum californici*) that went extinct during the captive breeding program of its host, the highly endangered California condor, *Gymnogyps californianus* (Pizzi 2009). This example demonstrates that we care more for the hosts than for their parasites, and that hosts may survive with the help of our conservation programs, while their parasites go extinct.

Emiliani (1993) and also Vredenburg et al. (2010) have postulated that, potentially, parasites and viruses could even drive their host to extinction during particular conditions. A strong decline has been observed in the total number of adult and subadult frogs in three metapopulations after the detection of amphibian chytrid fungus, *Batrachochytrium dendrobatidis* (Fig. 3.1; Vredenburg et al. 2010).

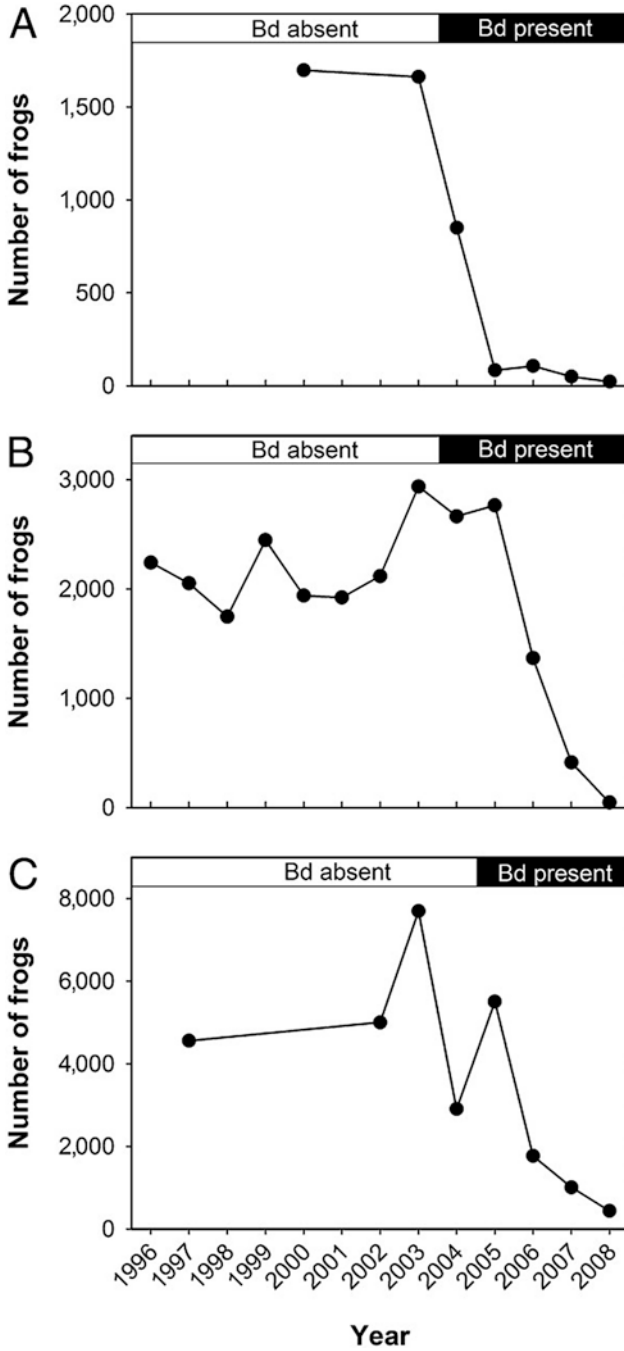


Fig. 3.1 The number of adult and subadult frogs in the different metapopulations before and after detection of Bd. Studied sites: Milestone Basin (a), Lake Basin (b), and Barrett Lake Basin (c). Image from Vredenburg et al. (2010)

However, the evidence for this dynamic is lacking from deep time, and the total exclusion of a species by another species is considered rare in nature (Vermeij 2004).

Traditionally, disease models suggest that parasites driving their host to extinction are highly unlikely to nearly impossible, and often the parasites go extinct before the hosts (De Castro and Bolker 2005). Some conditions, however, may allow for parasite-driven extinctions of a host, for example, a parasite which significantly reduces the reproductive capacity of infected hosts (Boots and Sasaki 2002), or the evolution of costly host defense that is limited by resource availability (Boots 2011).

However, resource availability and cost of resistance may actually play a more important role in the so-called killing the winner motifs (Våge et al. 2013; Våge et al. 2018). In this particular motif, the parasite (or virus) selectively attacks the “winner” (i.e., superior competitor or abundant) populations, preventing any single host from dominating a community and thereby promoting diversity (Thingstad and Lignell 1997). Empirical examples are few and far between, but they have been observed in land snails (Cunningham and Daszak 1998), and the red flour beetle, *Tribolium castaneum* (Rafaluk et al. 2015).

It is safe to assume that parasites and pathogens must have played an important and active role in past extinctions as evidenced by the global decline of amphibians caused by virulence, which is here defined as the parasite-induced reduction in host lifetime reproductive success (Herre 1993), and the emerging infectious disease chytridiomycosis, which was likely spread through the introduction of exotic species, and was enhanced by climate change and habitat destruction (Wake and Vredenburg 2008).

While plausible in theory, it is challenging to track host–parasite relationships empirically in the fossil record. It is reasonable to assume that parasitism has existed in one form or another since the beginning of early life. However, most parasites are small, and their soft bodies hardly fossilize (Leung 2017). However, some types of host–parasite collections allow us to track parasitic infestation in deep time (De Baets and Littlewood 2015; De Baets et al. 2021a, b; Wood et al. 2017). Parasites can leave traces in their host’s remains, which can be studied, as well as traces and (resistant) propagules in coprolites, the host’s fecal remains. Yet it is difficult to gather enough evidence to precisely identify host and parasite. In some cases, it is possible to track hosts and their associated parasites and/or characteristic pathologies, as with platyceratid gastropods in the fossil record (Baumiller and Gahn 2002).

3.3 Co-extinction

Global change is expected to typically lead to co-extinction of hosts and their specific parasites (Dunn et al. 2009; Carlson et al. 2017; but see Strona 2015). Some models that account for host-driven co-extinctions predict that up to 30% of the parasites go extinct in the form of such secondary extinctions (Fig. 3.2; Carlson et al. 2017). However, such a perspective would strongly depend on to what degree

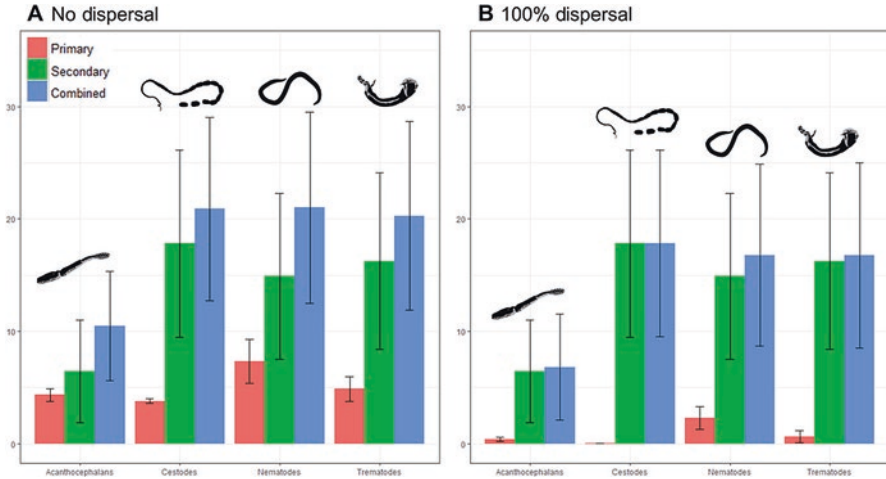


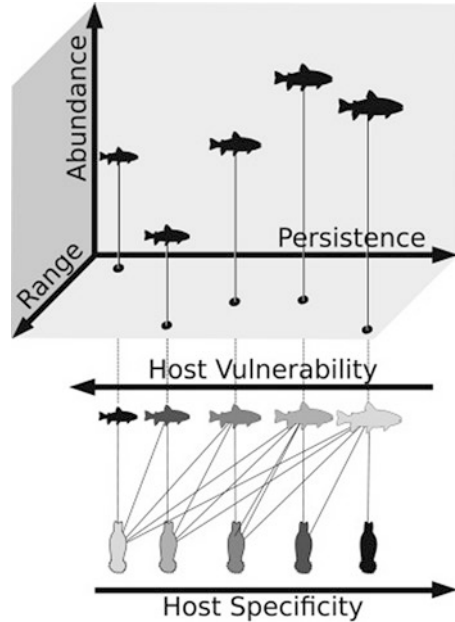
Fig. 3.2 Dispersal (a) and no dispersal (b) scenarios and the effects on the primary (due to climate change), secondary (co-extinction with hosts), and combined (total) extinction rates for major helminth clades. Figure from Carlson et al. (2017)

parasites specialize on niche specialists or top predators (Lafferty 2012). Models of parasite co-extinction with host decline differ from models that are based on host extinction alone. Such models of co-extinction only focus on the outcome after complete host extinction, but a species will generally show a significant decline in abundance and range size well before it finally goes extinct (Farrell et al. 2015).

Parasite survival has often been closely linked to the survival of its host; it can therefore be expected that when the host goes extinct, so will the parasite (Koh et al. 2004). Many tick species, for example, are endangered (Mihalca et al. 2011). Dunn et al. (2009) rightly posit that with current species' declines and extinctions, the most endangered species will actually be parasites and mutualists. This is because all species with a symbiotic lifestyle naturally depend on the availability of a host. Additionally, each host species likely harbors just as many, if not more, in symbiosis living species: from the mutualistic bacteria in our intestinal tract, the ones on our skin (e.g., lice and ticks), to parasitic flatworms and viruses.

Endoparasites are especially vulnerable to host-driven co-extinction (Carlson et al. 2017). Some are specialists, being highly host species specific, and thus potentially more prone to co-extinction, while others are generalists that may be able to adapt to the change in host availability. Co-extinction has a high chance to occur when the parasites are specialized on niche specialists or hosts higher on food chains (Lafferty 2012). However, this does not need to be the case when the host has a wide distribution and high abundance (Strona et al. 2013; Strona 2015). Highly specific parasites predominantly use low-vulnerability hosts (Fig. 3.3), and thereby reduce the risk of a co-extinction (Strona et al. 2013). Parasite and host co-extinction are often overlooked, but may have occurred throughout Earth's history, especially during mass extinction events (e.g., Seilacher 2007).

Fig. 3.3 3D representation of the relation between host vulnerability and host specificity. Image reproduced from Strona (2015)



Co-extinctions can be expected at a much broader scale and are not only restricted to host–parasite relationships, but concern all types of species that depend on one or more species in general: from the many forms of symbiosis to predators and their prey, to key species in an ecosystem. In a predator–prey relationship, for example, where the predator is highly specialized on just one type of prey, the predator has a high risk of co-extinction if it does not change its diet, but also herbivores with a very specific diet can fall victim to co-extinction (Labandeira et al. 2002).

Another, highly relevant, coevolutionary relationship is that between plants and pollinators. One-third of Europe’s crop plants depend on pollinators (Kearns and Inouye 1997). Local extinctions of bee populations and the parallel decline of insect-pollinated plants due to anthropogenically modified landscapes have been observed in several parts of Europe and are of much concern (Biesmeijer et al. 2006). Co-extinctions of plants and their pollinators is nothing new and has happened in the deep past as well (Bascompte and Jordano 2007). Thus, co-extinction may be the most common form of species loss (Koh et al. 2004; Dunn et al. 2009).

The disappearance of key species should thus be our main concern as that could result in a cascade of secondary extinctions and co-extinctions of the parasites and others that depend on them (Stork and Lyal 1993). The assumption has been that threatened hosts have relative fewer single-host parasites (Dunn et al. 2009; Lafferty 2012), but host extinction may vary across groups of hosts. Threatened ungulates were found to have a higher proportion of single-host parasites compared to non-threatened ungulates, a result related to a disproportionate decrease in richness of multi-host parasites, but among carnivores this relation does not exist (Farrell et al. 2015).

3.4 Dilution Effect

The current biodiversity loss and disease emergence have become two of the most challenging issues confronting science and society (Johnson et al. 2008). Different authors have found a strong correlation between parasite success in ecosystems and biodiversity of their ecological communities (Johnson et al. 2013; Lagrue and Poulin 2015), and many of them underline that the rapid loss of populations and biodiversity significantly increases overall disease risk. Studies concerning the causal relationship between biodiversity and disease emergence in a particular environment are focused on testing the “dilution effect” model (Ostfeld and Keesing 2000), which parasitologists refer to as the “decoy-effect” hypothesis (Combes and Moné 1987; Johnson and Thieltges 2010). According to these researchers, the mechanisms of the decoy effect, as observed in the case of high biodiversity in ecological communities, concern (1) the physical degeneration of parasite life history stages infecting a nontarget host, (2) the encounter reduction caused by infecting a nontarget host, and (3) the stimulation of defense mechanisms in nontarget hosts against the infectious stages of the parasite. Regardless of the mechanism, the nontarget host becomes a dead-end host, which is the real factor reducing the emergence of parasitic disease (Mehlhorn 2008). Essentially, this is in the same line of thought as Keesing et al. (2006), who propose that the term “dilution effect” should be interpreted as “the net effect of increased species diversity reducing disease risk.”

The dilution hypothesis has been in particular investigated for zoonotic diseases, like Lyme disease, but has also been reported to be more widespread (Civitello et al. 2015), and has even been reported for zooplankton (Hall et al. 2009). However, a meta-analysis of classical studies with new approaches has demonstrated that research is heavily biased towards studies presenting the dilution hypothesis, which yield biased results (Young et al. 2013). When correcting for this, a meta-analysis could not find strong evidence for the dilution hypothesis (Salkeld et al. 2013). Further analyses suggest that there is a slight publication bias towards negative relationships between biodiversity and disease risk (Fig. 3.4; Salkeld et al. 2013).

The generality of the dilution hypothesis is still debated and might be context dependent. It might particularly work on local scales, while large-scale analyses usually find a positive correlation between host and parasite diversity (Wood and Lafferty 2013). In that sense, it is mostly a matter of scale whether one observes a dilution effect or not (Fig. 3.5; Hopkins 2013), and may further depend on the characteristics of host communities (Halliday et al. 2020) and species interactions, such as predation (Su et al. 2020). Large-scale studies show mostly the traditional patterns, and this is potentially also the dominant relationship we might find on longer evolutionary patterns observed in the fossil record. Preliminary data compiled by Baumiller and Gahn (2002) on the prevalence of parasitic pathologies in marine invertebrates are very reminiscent of the traditional perception of metazoan diversity (the so-called Sepkoski (1981) curve). A similar positive relationship also exists between infested species of crinoids and crinoid diversity (Fig. 3.6; compare Baumiller and Gahn 2002; Baumiller et al. 2004). Irrespective of the presence of the

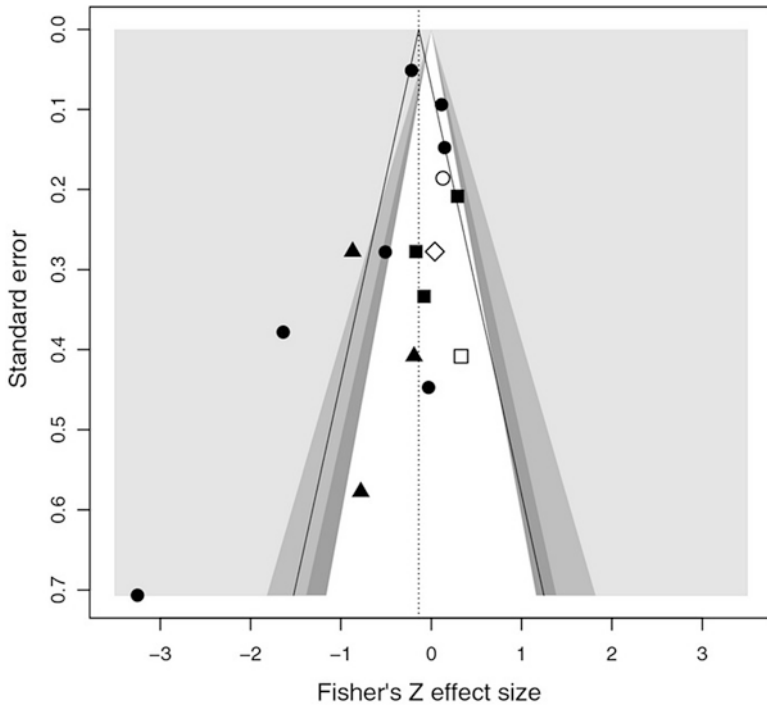


Fig. 3.4 Funnel plot of the relation between the Fisher's Z effect size and the standard error for studies on biodiversity and disease risk. Adapted from Salkeld et al. (2013)

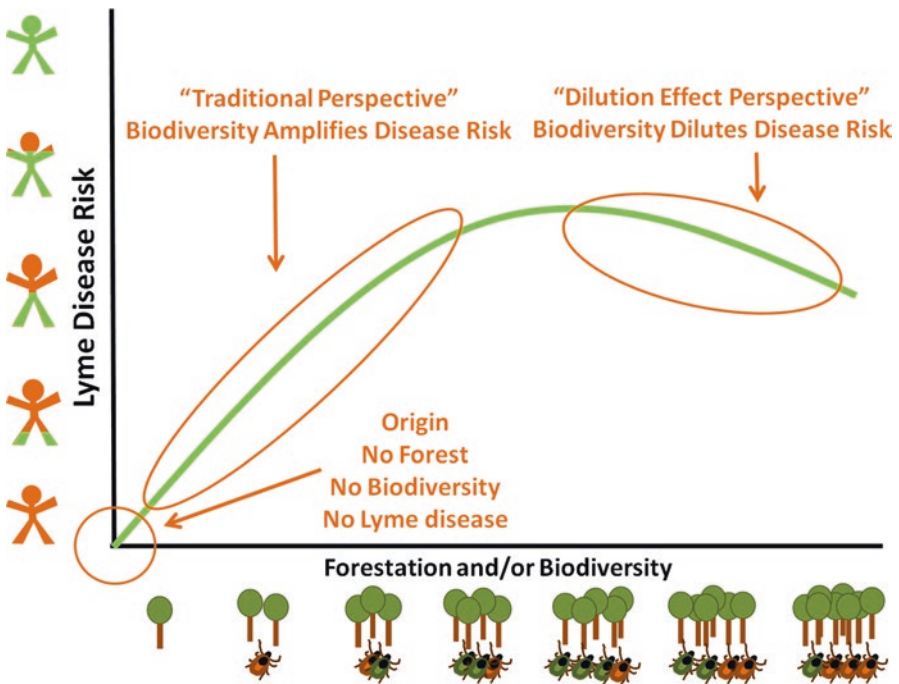


Fig. 3.5 As forest area increases, tick density will increase, but less fragmentation will lead to a decrease of infection prevalence. Lime green is used to indicate Lyme infection. Figure by Hopkins (2013)

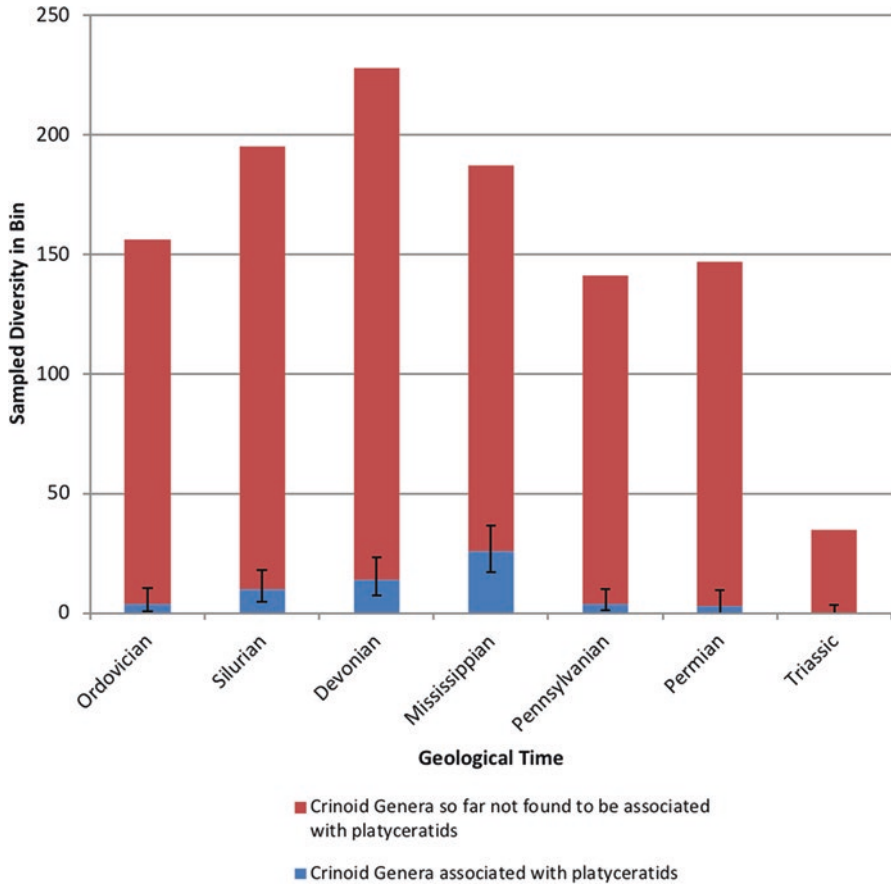


Fig. 3.6 Crinoid genera parasitized by platyceratids during the Paleozoic. Entire bars show the crinoid generic diversity based on sampled-in-bin diversity downloaded from the Paleobiology Database (PBDB, accessed in 14/11/2019). Co-occurrences of crinoids and platyceratids are shown in the solid blue part with error bars for 95% confidence levels. Based on data from Baumiller and Gahn (2002)

dilution effect or opposite patterns—this context dependency is still of great interest on larger timescales.

How can we extrapolate anything like the dilution effect to deep time, especially when there is no general consensus on the dilution effect occurring in the modern world? In fact, some workers even question how effective a dilution effect actually is in reducing disease risk. Critique on the dilution effect model concerns the issue that it may only work if the parasite is (more or less) a host specialist, and increasing host diversity will also increase infection prevalence; thus the addition of host species may actually increase parasite abundance (Randolph and Dobson 2012; Wood and Lafferty 2013). The mechanisms behind the dilution effect are complicated, even if some aspects of it might be applicable to the fossil record, e.g., trace fossils

that indicate a parasite with host preference, and high vs. low variation in potential hosts; but these findings will first and foremost indicate density dependence. Let us not forget that the dilution effect itself refers to the reduced *disease risk* for a vulnerable focal host at higher biodiversity due to the presence of more diluter species (Keesing et al. 2006). These diluter species can in several ways decrease transmission: they can affect focal host behavior, reduce focal host population density so it may become fragmented, or may feed on the disease vectors, e.g., the parasites (Keesing et al. 2006; Keesing et al. 2010). Thus, parasites do play a role here as a vector transmitting the disease (e.g., ticks spreading Lyme disease), but the transmitting-part is hardly observed in the fossil record, if at all. To actually observe this in the fossil record we would need to investigate three things: (1) an indication that the mortality of selected fossils was caused by disease, (2) the prevalence of parasite traces, and (3) whether 1 and 2 are related, or not. Morphological studies of fossil remains could be one approach, as disease may affect morphology as well as size. Infected organisms may compromise with reduced growth (Ruiz 1991), a phenomenon reminiscent of the “Lilliput effect,” which was observed in the context of mass extinctions (Urbanek 1993). However the opposite may also have happened in some cases, in the form of pathological gigantism, possibly caused by parasitic castration (Manger et al. 1999; but see De Baets et al. 2015). Modern studies show that skeletal pathologies observed in helminth-infested frogs are inversely correlated with survival (Johnson et al. 2011). The number of malformed amphibians has markedly risen (Fig. 3.7, Johnson and Chase 2004)—making it tempting to attribute this to anthropogenically induced factors (Johnson and Chase 2004).

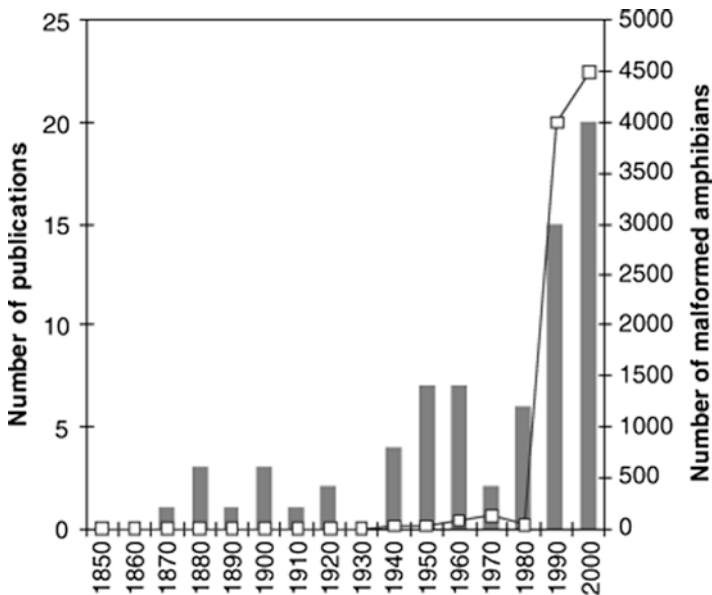


Fig. 3.7 The number of articles published on, and the number observed in (line), malformed amphibians in wild populations in North America (USA and Canada). Figure from Johnson and Chase (2004)

An observed dilution effect may well be a “chicken or the egg” causality dilemma, and in fact, there may be more a “killing the winner” dynamic going on, than an actual dilution effect. Indeed, it may be a matter of scale as some workers have pointed out, but it may also depend on the moment in (long-term) time. Much like the “Red Queen” dynamics in predator-prey interactions, populations of hosts will likely follow a sinusoidal pattern through time. In turn, Red Queen dynamics might be one mechanism that promotes host switching (Rabajante et al. 2016). A highly successful and abundant species may attract more parasites, and then decline because of parasite load, which then allows the host’s competitors to rise. As a result, parasite density may fall and/or the parasites may switch to a new host, which, in turn, would allow the original host population to strengthen and increase again. The presence of competition may in fact be beneficial to the host, driving rapid evolution of hosts with high phenotypic variation and may thereby “rescue” the host’s population densities despite larger epidemics (Strauss et al. 2017).

The notion that diverse communities inhibit the proliferation of parasites (Civitello et al. 2015) is in that sense incorrect, because in reality it is more likely that parasites inhibit the proliferation of any single (or multiple) host species. Much like the “paradox of the plankton,” which addresses the situation where a number of phytoplankton species are able to coexist in the same environment while competing for the same resources (Hutchinson 1961). Here, the parasites play an important role in facilitating diversity by allowing multiple species to exploit a certain niche (Våge et al. 2018). The deep-time fossil record could well play an important part here if only we could find an effective way to reconstruct these dynamics.

A complicating factor here might be that an increase of host pathologies in the fossil record could relate to an increase in parasitized specimens, but could also reflect an increase in abundance of hosts that can better cope with being parasitized, or at least with the developing pathologies. A positive relationship between sampling opportunity and finding pathologies might also affect such a relationship (cf.) as it has been suggested for predation prevalence (Huntley and Kowalewski 2007). However, by looking at larger samples of hosts from various localities and time intervals one could get an idea about the relationship between host population structure and prevalence of pathology changes through time. Such a larger sampling would also avoid the pitfalls of small sample size.

3.5 Host Switching

Host switching, e.g., parasites “jumping” from one species to the other, broadly speaking, occurs naturally as part of the life cycle of many species of parasites. On evolutionary timescales this usually refers to events where parasites switched host, which is usually inferred from coevolution patterns in phylogenies (Page 2003; Martínez-Aquino 2016). However, these phylogenies might be hard to resolve and not entirely equivalent (e.g., De Vienne et al. 2013; Poisot 2015). The biology of host switching revolves around three factors: (1) the rate of exposure of the new host

to the parasite, (2) the compatibility of the pathogen towards the new host, and (3) whether the pathogen is sufficiently transmissible between individuals within the new host population (Woolhouse et al. 2005).

Host switching has played an important role in the evolution of many parasite groups, but most of these host-switch events have been inferred solely from molecular phylogenies of extant taxa (e.g., Badets et al. 2011). Fossil parasites may help us to calibrate molecular clocks in such trees, as well as confirm past host-switching events (De Baets and Littlewood 2015; Leung 2017; Warnock and Engelstädter 2021). The fossil record has also revealed various combinations of parasite and host (e.g., arthropod and their pathogens) that have no extant equivalents (see De Baets and Littlewood 2015; Leung 2017, 2021 for reviews).

Ticks (order Ixodida) are known to have switched hosts many times during their evolution. It is therefore likely that host specificity is merely temporal and determined by biogeography and ecology (Klompen et al. 1996). One example is *Nuttalliella namaqua*, a monotypic tick species (the only representative of its genus). Phylogenetic analysis placed *N. namaqua* basal to the Ixodida, and can therefore be considered a “living fossil,” with its ancestors originating in the Late Carboniferous to Early Permian (Mans et al. 2011; Mans et al. 2012). These ancestors must have parasitized early reptiles and evidently changed host preference to mammals and lizards (Mans et al. 2014; but see Dunlop 2021).

Traces of host switches can also be identified via horizontal gene transfers (HGT) or horizontal transfers of retrotransposons (HTT) in the genomes of hosts and parasite species. For example, lymphatic filariasis and loiasis are two widespread human diseases caused by insect-borne filarial nematodes *Brugia spp.*, *Wuchereria bancrofti*, and *Loa loa*. These nematodes were likely endoparasites of tropical birds during the Oligocene/Miocene epochs (Suh et al. 2016; Suh 2021)—both of these groups were at least present at the same time in the same regions, as evidenced by amber records (Poinar et al. 2007; Poinar 2010, 2011a, b). Evidence was found that the genomes of these nematode species share the retrotransposon AviRTE with seven lineages of tropic birds, which must have come from two waves of horizontal gene transfer (Suh et al. 2016).

3.6 Parasites as Drivers and Regulators

Parasites have various ways to affect the lives of their host. They can, for example, deteriorate the host’s health, or influence the host’s reproductivity or even host behavior, as seen in rats and mice. The common brain parasite, *Toxoplasma gondii*, influences the behavior of rats and mice (the intermediate hosts) to become easier prey for cats (the target host). Infected intermediate hosts show more exploratory behavior and are less fearful of cats. Humans can also be infected, but normally serve as a dead end, although it is interesting to hypothesize how *T. gondii* may have influenced our behavior and culture in the long term, after centuries of exposure.

Climate change and biotic invasions of disease vectors promote the transfer of novel diseases and parasites to native species (Tylianakis et al. 2008). Higher

temperatures can both enhance host susceptibility to parasites and reduce host survival and fertility (Traill et al. 2010). The most optimal parasite and host phenotype, in terms of infectivity and resistance, respectively, vary from one time point to the next leading to fluctuating selection dynamics (Hall et al. 2011). In turn, this may result in fluctuating “Red Queen” coevolutionary interactions, where the parasite is locally adapted to infect sympatric host species, but has trouble infecting allopatric hosts (Brockhurst et al. 2014). In natural systems this leads to negative frequency-dependent selection, where the parasite follows the most common host over time (Decaestecker et al. 2007; Wolinska and Spaak 2009). Parasite diversity can then be maintained through negative frequency-dependent selection and multiple-niche polymorphism (Radolf and Samuels 2010; Strona et al. 2013; Strona 2015).

The occurrence of diverse natural populations of asexual organisms can be explained by the presence of parasites, which seem to play an important role in maintaining host genetic diversity (Turko et al. 2018). Sexual reproduction may exist for the same reason: instead of clonal reproduction where each generation is basically a copy of the previous, sexual reproduction creates diversity, which enables populations to cope with parasite infection (Jokela et al. 2009). Competition and defense could be central structuring factors in some microbial communities (Våge et al. 2018). This may also lead to diversification of both host and parasite, even in a homogenous environment, such as with the prey and predator leading to the paradox of the plankton (Hutchinson 1961). This process has also been termed “killing the winner,” where a parasite or virus prevents a susceptible competitive host (the winner) from monopolizing a limiting resource (Thingstad and Lignell 1997). This in turn allows the coexistence of resistant hosts, even when their defense is associated with a cost in the form of reduced competitiveness.

While competitive hosts tend to be infected by virulent specialists, less virulent generalists infect more hosts with higher resistance. Species that have found a way to reduce the costs of defense against parasitism without losing too much in competitive ability may reach the highest abundances (Våge et al. 2018), although other factors such as predation may still prevent this. The resistant hosts may be resource controlled, disappearing at low resource levels and dominating at high resource levels (Våge et al. 2018). Highest diversity would then be found around intermediate resource levels, which seems to be a general pattern found in both microbial and macroorganism ecosystems (Smith 2007). Lower amounts of remaining resources will likely slow down the Red Queen arms race based on the reduced probability that resistant hosts can successfully exploit them. Over geological time, these arms races may have played an important role in structuring the food webs with its major functional groups (Våge et al. 2018).

3.7 What Can the Fossil Record Tell Us?

Host–parasite associations in the fossil record can be studied by comparing trait variation through time of the host, and whether or not they show signs of having been parasitized. The host may have reached adulthood, but compensated with

decreased growth, abnormal growth response, or some visible pathology, for example (Hengsbach 1990; Rothschild and Martin 1993; Dittmar et al. 2012). Certain defensive traits may have come about as a response to parasitism as well, as is the case with the tubed crinoids that evolved in response to parasitization by platycerata (Baumiller and Gahn 2002). In some cases, targetting of parasites by predators might also induce changes in the hosts such as spinosity in crinoids as a defense mechanisms against predation on platyceratid gastropods (Syverson et al. 2018).

Traces of parasites have also been found in coprolites, in the form of tapeworm eggs in ancient shark coprolites, for example (Dentzien-Dias et al. 2013). New methods such as using the synchrotron phase-contrast microtomography and high-quality virtual 3D reconstructions of coprolite inclusions may reveal ancient trophic relations (Qvarnstrom et al. 2017). An integrated approach of scanning combined with dissolving coprolites might be the most effective way forward (but see Wood and Wilmschurst 2016). Each of these systems has its own challenges.

The direct fossil record of viruses and unicellular pathogens is very limited, making it hard-to-test hypotheses about their origins and coevolution directly (Hayward 2017; Leung 2017)—other than by attribution of changes in diversity and/or abundance without recorded environmental perturbations. Nonetheless, viruses, and in particular retroviruses, have been found to leave endogenous viral elements (EVEs) behind in the genomes of hosts and previous (ancestral) hosts (Katzourakis and Gifford 2010; Holmes 2011). These EVEs can be used to explore ancient viral evolution and trace their origins (Aiewsakun and Katzourakis 2015; Aiewsakun and Katzourakis 2017).

Some of the more spectacular data involve ancient DNA, or aDNA (Lafferty and Hopkins 2018). Ancient DNA (aDNA) extracted from moa coprolites found in New Zealand revealed that some species of parasites survived the extinction of their hosts (Table 3.1), and still exist today (Wood et al. 2013; Boast et al. 2018). This dataset could indicate that since the extinction of moas, as much as 19 species of parasites

Table 3.1 Chart showing the distribution of parasite taxa across extinct species of moa, modern birds, and mammals, and the total number that appears, persists, or disappears

	Taxon	L. Bush Moa	Giant Moa	Upland Moa	H.-footed Moa	Kakapo	Modern birds	Mammals	Appear	Persist	Disappear
Apicomplexa	Eimeriidae ^a	1	3	9	3	0	3	0	0	3	6
	<i>Eimeria</i>	0	0	0	0	0	0	1	1	0	0
	Sarcocystidae	1	0	1	1	1	0	0	0	0	1
Ciliophora	Balantiididae	0	0	0	0	0	4	0	4	0	0
	Nematoda	0	1	0	0	0	0	0	0	0	1
Nematoda	Heterakoidea	6	6	7	0	0	0	0	0	0	9
	Seuratidae	1	1	1	0	0	0	0	0	0	1
	Panagrolamoidea	0	0	0	0	0	0	2	2	0	0
	Strongylida	0	0	1	0	0	0	11	10	1	0
	Platyhelminthes	0	1	1	1	0	0	0	0	0	1

Color fill illustrates presence (green) and absence (red)

^aEimeriidae: excluding *Eimeria*

^bAscaridida: unidentified species, excluding Heterakoidea and Seuratidae. Based on data from Boast et al. (2018), *SI Appendix*, Fig. S20

went extinct, while as little as 3 species might have survived and up to 4 might have newly appeared in birds. For example, three upland moa's eimeriid coccidia are still found parasitizing extant kiwis. On the other hand, up to 9 heterakoid nematode species that had coevolved to specialize on moas have disappeared since their host's extinction. Moreover, the disappearance of moas led to the secondary extinction of their predator and a cascade of secondary extinctions of host-specific parasites of both predator and prey.

The extinctions might however be overestimated as in modern birds only one (captive) kiwi and two (non-native) ostriches could be investigated, while the number of survivors and newly appeared species may have been underestimated for the same reasons. This is also supported by the fact that only 18 species of parasites went extinct when modern mammal samples are also included, while as few as 4 and as many as 17 species might have survived or appeared in the modern fauna investigated. Even though this approach might be suitable for investigating extinctions during relatively recent extinction events, it is limited when going further back in time. It does demonstrate that a significant proportion of parasite species can disappear when their hosts go extinct.

In more ancient (lithified) coprolite samples, identifying the parasites using aDNA is limited (Chin 2021; De Baets et al. 2021a; Greenwalt et al. 2021). This means eggs themselves can rarely be assigned on the species or genus level, but are usually only attributable to higher taxonomic ranks. Also, precisely identifying the hosts is difficult—especially when found isolated from their producers. This is not an issue per se, as one can still study the diversity and relative abundance of propagules and their relationship/dominance in coprolites assignable to larger groups. However, further work (e.g., Camacho et al. 2018) is necessary to understand the abundance of propagules and their relationship with parasite abundance in hosts as well as loss through preparation. More important, although sampled, their record is still comparatively patchy. Coprolites need to be more systematically investigated for parasitic remains throughout the Phanerozoic (Chin 2021)—ideally covering major climatic and/or extinction events. As it is difficult to assign coprolites precisely to their hosts, their precise relationship with host species might be limited.

Although the precise identification of the culprit of a particular pathology will be a challenge, the host can, in most cases, be identified up to genus or species level depending on the state of preservation. Moreover, the expression of the disease (paleopathology) can be precisely measured, and its prevalence in populations can be quantified. Through cost-benefit analysis and population studies it is even possible to establish their impact on growth and fitness of particular host samples (Baumiller 2003; Huntley and Scarponi 2012; Baumiller and Gahn 2018; Klompmaker et al. 2021; Zhang et al. 2020). Such an approach is necessary to understand the negative impact of parasites on their hosts. If we have densely spaced samples through time and/or space, we could even track how disease or infestation prevalence relates to characteristics of the hosts, such as abundance, evolutionary persistence, geographic range, mode of life, or degree of specialization. Particularly interesting systems are those that have modern analogues, like the isopod swellings in decapods (Klompmaker et al. 2014, 2021; Robins and Klompmaker 2019), and

trematode-induced traces in bivalves (Ruiz and Lindberg 1989; Huntley and De Baets 2015; Rogers et al. 2018; Huntley et al. 2021), which can at least be traced back to the Mesozoic. However, such studies are even more important when parasites have a higher preservation potential—like platyceratid gastropods shown to be parasitic on crinoids (Gahn and Baumiller 2003).

Considering that remains of shelled invertebrate hosts are more widespread—they do allow us to investigate the precise contribution of shifts in pathology prevalence, environmental perturbations, diversity fluctuations and mass extinctions. More importantly, the relationship between specificity, or the prevalence of pathologies and host characteristics, can be investigated. This could be used to investigate whether the perception holds true that parasites tend to specialize on a specific host or rather on a number of resistant hosts. A better idea of this on longer timescales would also be crucial to understand the future of parasite diversity, as well as disease prevalence, in their hosts. Climate change for example, and in particular sea-level rise, has been linked to an increased prevalence of trematode infestations in bivalves (Huntley et al. 2014; Scarponi et al. 2017).

Cost-benefit analysis and modelling can help to establish the nature of this association (Baumiller 2003; Baumiller and Gahn 2018). Furthermore, the relationship between diseased species (specificity) and infested individuals within samples (prevalence) could be investigated. This, in principle, allows us to investigate how these properties relate to host persistence, abundance, and/or geographic range.

In the case of amber inclusions, novel techniques, such as phase-contrast synchrotron X-ray tomography (Dunlop et al. 2011; Dunlop et al. 2016), may allow the precise identification of the host, and the identification of the parasite to genus, or (at least) family level, might well be within grasp. However, more quantitative studies of amber inclusions are necessary to establish the prevalence of such associations. Due to a limited number of characters, it is still not that straightforward without knowledge about the life cycle of modern relatives and their host associations, to precisely identify particular parasites with limited external characters such as nematodes (Poinar 2011a, 2011b). Research has, at least, revealed potential new extinct, or rare, host associations (Peñalver et al. 2017), as well as tracked ancient host associations back in time (Haug et al. 2021; Labandeira and Li 2021). Another factor which needs to be considered when using amber deposits remains the debate concerning the age of many deposits, their patchiness in time, and the selectivity of trapping and preserving hosts (McCoy et al. 2018; Solórzano-Kraemer et al. 2018; De Baets et al. 2021a).

Despite these limitations, we are convinced that studying these model systems would advance our understanding of host–parasite evolution in deep time, as well as further constrain the modern baseline. Amber inclusions would be particularly good to understand the evolution and extinction of arthropod parasites and vectors, while invertebrate pathologies and propagules in vertebrate coprolites might be the only way to cover the impact of mass extinctions on parasitic disease and their link with host diversity. Before we can fully exploit the fossil record for this purpose, more data needs to be collected still, before such studies become feasible. Currently, most of these systems are incompletely studied—larger samples in particular are rare, and reports of prevalence of certain pathogens or impact on host population are still

limited or restricted to particular time intervals (e.g., De Baets et al. 2021b). Systematic screening of museum collections yielding large samples of well-determined hosts from particular localities might help considerably in such an endeavor (Harmon et al. 2019), as well as considering these constraints when sampling and describing new host remains.

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